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# $PM_{10}$ and the Air Quality Stress Index in an industrial city of Eastern Mediterranean

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The contribution of  $PM_{10}$  to the daily Air Quality Stress Index (AQSI) was examined in a heavily affected industrial city of Eastern Mediterranean (Aspropyrgos, Greece). For this purpose, hourly concentration values of four pollutants (SO<sub>2</sub>, NO<sub>2</sub>, O<sub>3</sub>, and PM<sub>10</sub>) were analyzed between 2012 and 2020, revealing that the main contributor to AQSI levels came from PM<sub>10</sub> (between 17% and 90% to the daily AQSI), with a moderate annual variability and a spring peak. Excluding PM<sub>10</sub>, the AQSI always remained below the threshold of 0.8. To identify the atmospheric source of PM<sub>10</sub> peaks, the Flextra – Air mass trajectories model was applied to 47 cases in which the upper threshold was exceeded. The results of the model show that dust transport episodes, mainly from Sahara Desert, contribute to the daily levels of PM<sub>10</sub> and, accordingly, to AQSI.

Keywords: PM<sub>10</sub>, Air Quality Stress Index, dust transport episodes

## 1. Introduction

The first and main industrial area of Greece is located in West Attica Prefecture, west-northwest of Athens, the capital city of Greece, which hosts 40% of the country's population. The prefecture has the industrialized Thriasio Plain as the core area located 20 km northwest of Athens, including three municipalities, namely Aspropyrgos, Elefsis and Mandra, along with the community of Magoula (Figs. 1a and 1b). The topography is mostly flat, with a gentle slope towards the sea not exceeding 3%. This basin encompasses an area of 500 km<sup>2</sup> and houses a total population of 120,000 inhabitants (2011 census, ELSTAT, 2011). The presence of several mountains, close to the sea, the local climatic conditions (Mavrakis et al., 2015b), and the unfavourable topography, produce



**Figure 1.** *a*) Map of Attica prefecture, where Thriasio Plain and Athens basins and the air quality station of Alonistra (AL) – Municipality of Aspropyrgos, are shown. *b*) Urban Atlas – European Environment Agency (EEA, 2016) land-use map: industrial/commercial areas are displayed in purple, urban areas in red, rural areas in yellow, bare land in beige, vegetated land is displayed with different shades of green and light gray for the legislated municipalities plan.

local regressive atmospheric circulation patterns (Lykoudis et al., 2008) that inhibit the capability of atmospheric self-cleaning through dispersion and transport mechanisms (Mavrakis et al., 2008). Temperature inversion heights are especially low during the cold seasons of the year, frequently being lower that the surrounding hills and comparable to the heights of the highest chimneys characteristics of the large industrial compounds in the area. Even more, the sea shore is occupied by various industries, as well as other (authorized and informal) activities, which actually act as a wall between the shallow Elefsis Gulf and the inside plain (Mavrakis et al., 2008). This means that air pollutants are trapped within a shallow layer, resulting to high (daily and hourly) air pollutants' concentrations. The Thriasio Plain exhibits the higher: (a) industrial activity concentration, (b) fuel consumption and (c) pollution, related to the production processes in Greece (Abatzoglou et al., 1996; Lykoudis et al., 2008; Mavrakis et al., 2008). Mavrakis et al. (2016) quantified the industrial asset of the area in (at least) 6,500 large and small industrial plants, including all kinds of industries, such as oil refineries, chemical and steel production, metal and non-metal minerals' processing, cement industries and logistics (Fig. 1b). Apart from the industrial activities, the area is crossed by three freeways (two national roads and, since 2003, a large part of the urban freeway "Attiki Odos"), as well as by two railroad lines (the national railway line and, since 2005, the Athens' suburban railway). Furthermore, nine large quarries are actually under operation, and the 13 docks of Elefsis harbour accommodate 5,500 ships per year, with a total cargo load 2.5 times larger than that handled by the Piraeus harbour (the major one of Greece).

The municipality of Aspropyrgos houses about 35,000 inhabitants and receives the largest number of all kinds of established activities (2,700 out of a total of 6,500; Fig. 1b), around the municipal centre (Di Feliciantonio et al., 2018), thanks to the large availability of flat and fertile free land (both public and private) around the urban core of the city, contrary to the surrounding municipalities (Mavrakis et al., 2016).

For the air quality of the nearby Athens, the Thriasio Plain has been frequently regarded as a pollution source. However, a relatively limited number of studies considering air pollution issues, although pollution levels have been identified as serious, since the early 1990 (Abatzoglou et al., 1996; Kassomenos et al., 2004; Lykoudis et al., 2008; Mavrakis et al., 2008; Toumpos et al., 2017). Koukoulakis et al., (2019) and Kanellopoulos et al. (2021) have given special attention to primary and secondary organic aerosol and to trace elements bound to PM<sub>10</sub> in the area as well as to their possible health implications.

The present study examines trends in air quality over the city of Aspropyrgos, which is located at the centre of Thriasio Plain, using a composite index derived from the levels of four pollutants recorded in the area (namely  $SO_2$ ,  $NO_2$ ,  $O_3$ , and  $PM_{10}$ ), by using the complete hourly and daily data set, collected by the local air-quality monitoring station during the time period 2012–2020. The contribution of  $PM_{10}$  to the daily Air Quality Stress Index (AQSI) was finally examined. The Flextra – Air mass trajectories model was used to identify the atmospheric source of elevated  $PM_{10}$  peaks in all cases when the AQSI upper thresholds were exceeded.

#### 2. Methods

The data used in this study were provided by Alonistra (AL) air pollution monitoring station, located in Aspropyrgos town (500 m south of the city centre and 1500 m from the sea shore) and operated by the Bureau of Environment, Municipality of Aspropyrgos (BEMA). Hourly values of air temperature (Ta, °C), relative humidity (RH, %) and concentrations (µg/m<sup>3</sup>) of sulphur dioxide (SO<sub>2</sub>), nitrogen dioxide (NO<sub>2</sub>), ozone (O<sub>3</sub>) and particulate matter less than 10 µm in diameter (PM<sub>10</sub>), as recorded for the 2012–2020 time periods, were collected and analyzed here.

In European Union (EU), although member states are subject to the same directives in respect of air quality, the use of a common Air Quality Index (AQI) has not been adopted, with most countries applies their own AQI. This is because AQIs were developed in most of countries and thus have been adapted to technical requirements, quality standards and Knowledge needs of the investigated area (Karavas et al., 2021).

For the purpose of this study, air quality was assessed using the daily Air Quality Stress Index (AQSI). This methodology was firstly applied in SW Germany by Mayer et al. (2004) and also in Athens by Katsoulis and Kassomenos (2004) and Kassomenos et al. (2012). Air Quality Stress Index give an overall assessment of air quality, including the synergetic effects of the recorded air pollutants, is based on an arithmetic summation of relative concentrations of air pollutants and can be considered a summary assessment of the ambient air pollution. According to Mayer et al. (2004), Katsoulis and Kassomenos (2004) and Kassomenos et al. (2012) the typical structure is given by the formula:

$$AQSI = \frac{1}{n} \sum_{i=1}^{n} \frac{Ci - average - daily - values}{MI - 24h - values_i},$$
(1)

where *n* is the number of air pollutants, Ci is the time-specific concentration, and MI - 24h value indicates the threshold of air pollutant concentration, Based on the threshold values, the equation (1) can be rewritten as follows:

$$AQSI = \frac{1}{4} \left[ \frac{C(SO_2)}{350} + \frac{C(NO_2)}{200} + \frac{C(O_3)}{180} + \frac{C(PM_{10})}{50} \right],$$
 (2)

where *C* indicates the concentration of the respective pollutant expressed in  $\mu g/m^3$ . *C*(SO<sub>2</sub>) and *C*(NO<sub>2</sub>), are the daily maximum1-h concentrations; *C*(O<sub>3</sub>) is the daily maximum 8-h running mean concentration and *C*(PM<sub>10</sub>) is the daily

mean concentration value. The denominators stand for the upper limit values set by the EU directives (*i.e.*, hourly SO<sub>2</sub>: 350 µg/m<sup>3</sup>; hourly NO<sub>2</sub>: 200 µg/m<sup>3</sup>; mean 8hourly O<sub>3</sub>: 180 µg/m<sup>3</sup>; daily PM<sub>10</sub>: 50 µg/m<sup>3</sup>).

Katsoulis and Kassomenos (2004) gave the following description of stress categories scale for daily AQSI: (i) Class I (AQSI < 0.2): "Very Low" (ii) Class II (0.2 < AQSI < 0.4): "Low"; (iii) Class III (0.4 < AQSI < 0.6): "Moderate"; (iv) Class IV (0.6 < AQSI < 0.8): "Distinct"; (v) Class V (0.8 < AQSI < 1): "Strong" and, finally, (vi) Class VI (AQSI > 1): "Extreme".

Pearson correlation coefficients were used to investigate the pair-wise relationship between air temperature, relative air humidity, air pollutant concentrations, the contribution of each pollutant in the AQSI and daily AQSI values. The confidence level was set at p = 0.05. Since the scope of this paper is to examine the contribution of PM<sub>10</sub> to the daily AQSI and to identify non-local dust sources, when the upper thresholds of AQSI were exceeded, then the 'Flextra' (Air mass trajectories) model was used to calculate air masses back trajectories, their origin and the type of weather systems occurred. FLEXTRA is an atmospheric trajectory model. It can compute both forward and backward trajectories and can be driven by meteorological input data from a variety of global and regional models including ECMWF analyses and forecasts (Stohl, 1998).

## 3. Results and discussion

All the available air pollutants data per day were used for the calculation of the daily AQSI and descriptive statistics were provided separately for each pollutant and the composite index (Tab. 1). Sulphur dioxide (SO<sub>2</sub>) concentrations, both average and maximum daily or hourly values, have been greatly reduced since the 1980s over the entire study, and are now well below the EU Directive threshold. There are very few and short terms exceeding of thresholds, so a limited contribution of this air pollutant to daily AQSI values was observed.

Nitrogen dioxide (NO<sub>2</sub>) concentrations remained quite stable throughout the study period, being systematically under the threshold defined by CEC Directive 99/30 – both as average values and extremes. Although the meteorological conditions in the area – especially during morning hours (morning calms, weak sea breeze) – led to high NO<sub>2</sub> concentrations, only rare exceeding of the normative threshold was recorded.

The study area is also featuring (HNMS, 2021) high insolation (about 2,800 hours of sunshine per year) and considerable air temperature (mean value for summer months: 27.8 °C), stimulating formation and consolidation of photochemical pollution. Ozone ( $O_3$ ), due to its photochemical nature, shows a marked seasonal trend, depending on solar irradiation. During summer months, the 8-hours moving average exceeded the EU threshold. Ozone, as linked to specific sources (Lykoudis et al., 2008).

		2012	2013	2014	2015	2016	2017	2018	2019	2020
$SO_2$	Average	8	6	6	8	7	7	6	6	10
	$\operatorname{StDev}$	10	5	6	8	8	<b>5</b>	5	4	6
	Min	0	0	1	1	1	0	0	1	2
	Max	62	28	55	57	57	30	28	21	52
$NO_2$	Average	35	48	36	41	51	39	46	47	41
	$\operatorname{StDev}$	16	22	14	17	18	14	15	16	15
	Min	3	8	5	5	16	8	13	14	10
	Max	81	140	92	90	110	77	94	101	92
$O_3$	Average	36	41	50	54	42	43	44	33	46
	$\operatorname{StDev}$	20	24	23	23	23	20	21	20	22
	Min	3	1	4	11	8	<b>5</b>	2	2	2
	Max	91	102	100	114	240	94	93	84	91
$\mathrm{PM}_{10}$	Average	58	54	52	41	47	42	49	49	42
	$\operatorname{StDev}$	20	21	51	16	23	17	25	20	18
	Min	24	11	9	15	14	8	11	8	3
	Max	113	154	726	124	283	112	310	123	115
AQSI	Average	0.32	0.44	0.42	0.42	0.45	0.41	0.44	0.48	0.41
	$\operatorname{StDev}$	0.14	0.16	0.25	0.12	0.15	0.12	0.15	0.17	0.13
	Min	0.10	0.06	0.02	0.14	0.16	0.18	0.14	0.09	0.12
	Max	0.81	1.19	3.90	0.84	1.70	0.89	1.78	1.25	0.94
AQ	$SI \ge 0.8$	1	8	6	1	5	1	5	16	4
A	$QSI \ge 1$	0	2	3	0	2	0	2	2	0

Table 1. Descriptive statistics of daily values of the air pollutants, the Air Quality Stress Index (AQSI), and number of exceedings, distinguishing two thresholds (0.8 and 1) in Aspropyrgos, by year.

 $PM_{10}$  concentrations in Aspropyrgos represent the most important pollutant in the area, maintaining very high for both daily and hourly values (Tab. 1) and the respective thresholds were frequently exceeded. According to earlier studies (Abatzoglou et al., 1996; Razos and Christides, 2010),  $PM_{10}$  corresponds to 80% of the Total Suspended Particulate (TSP) in the area. The relative and cumulative frequency [%] of  $PM_{10}$  classes [µg/m<sup>3</sup>] during 2012–2020 in Aspropyrgos was shown in Fig. 2. Only 43% of the recorded PM10 daily values were below the EU threshold (50 µg/m<sup>3</sup>). All the remaining records exceeded largely this threshold. Although earlier studies (Flocas et al., 2009; Dimitriou and Kassomenos, 2017; Matthaios et al., 2017) have reported similar cases from other sites in the Eastern Mediterranean basin, a persistence over time of particularly high values was observed in the study area.

Pearson correlation coefficients (Tab. 2) were calculated for the association between air temperature ( $T_a$ ), relative humidity (RH), air pollutant concentrations (SO<sub>2</sub>, NO<sub>2</sub>, O<sub>3</sub>, PM<sub>10</sub>), considering both average and maximum values, the contribution of each pollutant to daily AQSI (AQ\_xx), as well as and the AQSI



Figure 2. Relative and cumulative frequency [%] of PM10 categories  $[\mu g/m^3]$  of recorded values, during the 2012–2020 measurement campaign in the municipality of Aspropyrgos.

itself. Positive and statistically significant coefficients were shown in bold and negative, statistically significant, coefficients were marked with italics and light grey. An intense correlation between ozone and air temperature was demonstrated, together with the strong dependence of the AQSI from  $PM_{10}$ . The esti-



Figure 3. Variation of daily Air Quality Stress Index (AQSI) values. Thresholds of the index are marked with green, yellow, orange, and red lines.

Table 2. Pea centrations ( and the AQS italics and la particulate <i>n</i> contribution	rson cor SO <sub>2</sub> , NC SI. Positi ight grey natter le of PM <sub>10</sub>	relation $D_{2}, O_{3}, PI$ ive and i V (Abbreu ss than .)	coefficien $M_{10}$ ) –bot statisticc viations: 10 $\mu$ m in	nts inves h avera <u>e</u> ully signi Ta, air v diamete	tigating se and m ificant cc tempera er; AQ_S	the assoc aximum efficient: ture; RH ture; Contr	iation b values, s are shc , relativ ibution	etween c the conti twn in b e humia of SOs: 2	ir tempe ribution old; nego ity; SO <sub>2</sub> AQ_NO2	rature (1 of each p ative, sta sulphun contribu	fa), rela ollutant tristicall, dioxide tion of I	tive air h to daily y signifu y NO <sub>2</sub> , AQ	umidity Air Quc cant coef itrogen 	(RH), a dity Stre ficients o dioxide; ribution	ir pollut ss Index are shaa O <sub>3</sub> ozon of O <sub>3</sub> : A	ant con- : $(AQSI)$ led with e; $PM_{10}$ $Q_{-}PM_{10}$
	νвТ	xsmT	vвНЯ	хвшНЯ	VE2OS	$x \text{sm}_2 \text{OS}$	vb2ON	xsm <sub>2</sub> ON	$v_{B_g}O$	${\tt xsm}_{\it g}{\rm O}$	$\rm PM_{108V}$	X8m01MA	202_0A	20N_QA	8 <b>0_</b> 03	AQ_PM10
Tmax	0.99															
RHav	-0.62	-0.64														
RHmax	-0.60	-0.60	0.92													
$\mathrm{SO}_2\mathrm{av}$	-0.01	0.04	0.02	0.02												
$\mathrm{SO}_2\mathrm{max}$	-0.04	0.01	0.02	0.03	0.82											
$\mathrm{NO}_2\mathrm{av}$	-0.06	-0.01	0.18	0.18	0.31	0.27										
$\mathrm{NO}_2\mathrm{max}$	0.01	0.07	0.07	0.10	0.30	0.26	0.86									
$O_{3}av$	0.61	0.59	-0.67	-0.63	-0.15	-0.11	-0.51	-0.31								
$O_3max$	0.59	0.61	-0.54	-0.49	0.05	0.07	-0.19	0.00	0.82							
$\mathrm{PM}_{10}\mathrm{av}$	-0.03	0.00	0.19	0.19	0.22	0.20	0.39	0.33	-0.33	-0.12						
$\mathrm{PM}_{10}\mathrm{max}$	-0.02	0.01	0.19	0.22	0.17	0.14	0.31	0.30	-0.27	-0.08	0.79					
$AQSI_SO_2$	-0.02	-0.02	0.00	0.00	0.23	0.44	0.02	0.01	-0.01	-0.01	0.00	-0.01				
$AQSI_NO_2$	-0.02	0.02	0.06	0.07	0.25	0.22	0.64	0.75	-0.21	0.00	0.26	0.24	0.03			
$AQSI_0_3$	0.40	0.41	-0.37	-0.35	0.06	0.06	-0.10	0.01	0.48	0.54	-0.03	-0.05	-0.01	-0.01		
$\rm AQSI_{-}PM_{10}$	0.04	0.06	0.07	0.08	0.18	0.14	0.30	0.25	-0.23	-0.07	0.85	0.66	-0.01	0.18	-0.01	
AQSI	0.14	0.18	0.03	0.05	0.36	0.36	0.51	0.54	-0.17	0.11	0.92	0.73	0.06	0.42	0.17	0.79

mated correlation coefficients with AQSI was 0.72, 0.79, and 0.92, respectively for  $AQ_PM_{10}$ ,  $PM_{10}max$ , and  $PM_{10}av$ .

The diurnal variation of the daily AQSI was illustrated together with air quality thresholds (classes, see above) using colored lines (Fig. 3). Most of the recorded values was classified as "Moderate" and "Strong". Since a basic assumption of our study is that  $PM_{10}$  concentrations play a role in the annual variation



**Figure 4.** *a*) Month means of the daily Air Quality Stress Index with (black bars) and without (light gray bars) the contribution of  $PM_{10}$  air pollutant to index values. *b*) Relative Frequency of values with (black bars) and without (light gray bars) the contribution of  $PM_{10}$  air pollutant to AQSI classes, during the 2012–2020 measurement campaign in the municipality of Aspropyrgos.

of the index, the AQSI was recalculated excluding the  $PM_{10}$  (AQSI- $PM_{10}$ ) values. According to the results shown in Fig. 4, the AQSI- $PM_{10}$  had a clear annual variation, with a peak during summer months and a low during winter months. On the contrary, AQSI showed a weak annual variation, with slightly higher values during spring months. The contribution of  $PM_{10}$  to the daily AQSI range was estimated in a range between 17% and 90%. In agreement with other studies (Koukoulakis et al., 2019; Kanellopoulos et al. 2021), it was documented how  $PM_{10}$  was documented to have a daily impact to local air quality.

As illustrated in Fig. 3, the 0.8 EU threshold was overpassed frequently, delineating a strong health risk. A total of 47 exceeding events was identified, and 17 events exceeded a higher threshold (1), with a peak of 3.9.

Air masses back trajectories (Fig. 5) were generated using the Flextra model, which use meteorological data provided by ECMWF (European Centre for Medium Range Weather Forecast; www.nilu.no/trajectories; Stohl, 1998).

Results for 45 events from the 47 cases described above suggest that air masses transportation originated mainly from Sahara Desert and usually they circulated above central Mediterranean Sea and Greece, in agreement with previous studies (Aleksandropoulou and Lazaridis, 2013; Nastos, 2012). Two major types of circulation were identified (cyclonic and anti-cyclonic, and variations concerning the strengths of each observed system were frequently observed (Flocas et al., 2009; Nastos et al., 2011; Khomsi et al., 2020). Air masses above Mediterranean Sea usually transport significant amounts of Sahara dust and, at the same time, they enrich hot dry air with humidity (Kassomenos et al., 1998; Dimitriou and Kassomenos, 2017), having in turn a significant impact on human perception of weather conditions (Pantavou and Mavrakis, 2015; Pantavou et al., 2018; Cori et al., 2020). World Meteorological Organization (WMO, 2021) mentioned that dust transport episodes can be characterized as a hazard, for which official guidelines for preparation and mitigation of such events are largely unavailable (Palmos et al., 2021). Also, during the last few years, early-summer heat waves occurred in Greece, being accompanied with intense dust transport episodes. Those weather condition downgraded air quality (Trianti et al., 2017; Mavrakis et al., 2021) and alarmed authorities.

Air masses back trajectories were shown in correspondence with the three highest values of AQSI (a to c) and the main source areas were shown in Figs. 5d to 5f. The highest value of AQSI was recorded in 2014, June, the  $12^{\text{th}}$ , reaching 3.90 (Categorie 'Extreme – VI'). During all the day, the average and maximum recorded values were the same (726 µg/m<sup>3</sup>), and the episode was associated with anti-cyclonic conditions (a high pressure system dominated in Central Mediterranean Sea). This was one of the longer lasting dust transport episode recorded during measurement campaigns. During this episode lasting from May  $25^{\text{th}}$  to June  $12^{\text{th}}$ , 2014, two times maximum hourly values of the PM<sub>10</sub> were recorded equal to  $921 \text{ µg/m}^3$ . The second extreme case (AQSI = 1.78) was recorded in 2018, March, the  $26^{\text{th}}$  (Fig. 5b), and it was associated with cyclonic conditions

(a deep low pressure system located west of Greece in Central Mediterranean Sea).  $PM_{10}$  average daily value was 310 µg/m<sup>3</sup> and the maximum hourly value was 525 µg/m<sup>3</sup>. The third episode (AQSI = 1.70) was recorded in 2016, March, the 23<sup>th</sup> (Fig. 5c) and it was associated with cyclonic conditions (a shallow low



**Figure 5.** Air masses back trajectories for (*i*) the highest values of AQSI: *a*) 12/06/2014 at 12:00 UTC; *b*). 23/03/2016 at 12:00 UTC; *c*) 26/03/2018 at 00:00 UTC, and for (*ii*) transport area sources: *d*) (Sahara desert) 05/02/2021 at 18:00 UTC; *e*) (Central Asia desert) 10/06/2014 at 06:00 UTC; *f*) (Midle East desert) 19/10/2018 at 06:00 UTC.

pressure system moving across North Africa seashore). PM10 average daily value was  $283 \mu \text{g/m}^3$ , while the maximum hourly value approached 904  $\mu \text{g/m}^3$ .

As far as transport area sources, North Africa region/Sahara desert was the mean source of dust transport in 45 out of 47 cases (a typical example was the case of 2021, February, the 5<sup>th</sup>, Fig. 5d). Air mass back trajectories identified one case of dust transport from Central Asia deserts in 2014, June, the 10<sup>th</sup> (Fig. 5e) and one case of dust transport from Middle East deserts in 2018, October, the 19th (Fig. 5f).

#### 4. Discussion

Our study has investigated the contribution of  $PM_{10}$  air pollutant to daily Air Quality Stress Index in the socio-environmentally stressed industrial city of Aspropyrgos located in Westerm Attica, Greece, regarded as the most powerful pollution source of Athens. The results indicate that more than 55% of  $PM_{10}$ values exceeded the EU threshold. This made  $PM_{10}$  concentrations the main contributor to AQSI levels. Excluding  $PM_{10}$ , the daily values of AQSI remained always below the normative threshold and showed a clear annual variability. Including  $PM_{10}$  values, the AQSI showed a weak annual variability with a spring peak. To identify the source of  $PM_{10}$  values under exceeding of the AQSI upper threshold, the Flextra model was used. Results show that dust transport episodes mainly from Sahara desert, is a natural contributor to the daily levels of  $PM_{10}$ and AQSI levels and those episodes reach their peak during spring and autumn months.

The results from the present study are in agreement with previous studies (Cori et al., 2020) and could be used in environmental awareness campaigns informing population about the negative consequences of  $PM_{10}$  / dust transport episodes and alerting citizens toward adoption of individual prevention measures. The empirical results of this study indicate that policy makers have to consider additional prevention and mitigation plans to meet new health risks for general population regarding  $PM_{10}$  concentrations. This could also be a planning and risk management tool for policy and decision making authorities to chart their long-term future actions with a better understanding of potentially unfolding scenarios and their impacts.

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#### SAŽETAK

# PM<sub>10</sub> i indeks pritiska onečišćenjem zraka u industrijskom gradu na istočnom Sredozemlju

#### Eleni Verouti, Dimitrios Gavathas i Anastasios Mavrakis

Ispitan je doprinos  $PM_{10}$  dnevnom indeksu pritiska onečišćenjem zraka (AQSI) u industrijskom gradu na istočnom Sredozemlju (Aspropyrgos, Grčka). S tim ciljem analizirane su satne vrijednosti koncentracija četiriju polutanata (SO<sub>2</sub>, NO<sub>2</sub>, O<sub>3</sub> i PM<sub>10</sub>) za razdoblje od 2012. do 2020. Rezultati su pokazali da razinama AQSI najviše doprinose  $PM_{10}$  (između 17% i 90% za dnevne vrijednosti AQSI) uz umjerenu godišnju varijabilnost i proljetni maksimum. Ako se iz analize izuzmu  $PM_{10}$ , vrijednosti AQSI uvijek ostaju ispod praga od 0.8. Kako bi se identificirali izvori epizoda  $PM_{10}$ , na 47 slučajeva u kojima je prekoračen gornji prag, primijenjen je model putanja česti zraka – Flextra. Rezultati modela pokazuju da epizode transporta prašine, prvenstveno iz Sahare, doprinose dnevnim razinama  $PM_{10}$  i, u skladu s tim, razinama AQSI.

Ključne riječi: PM<sub>10</sub>, indeks pritiska onečišćenjem zraka, epizode transporta prašine

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